Contribution of the surface diffusion on mass transfer of Ba²⁺ in granite matrix

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The effective diffusivity (D_e) and the distribution ratio (K_d) of Ba^{2^+} in Inada granite have been determined by the through-diffusion method. Stable $BaCl_2$ solutions of the equal concentration were placed in two reservoirs across a granite sample and the self-diffusion of Ba was observed using radioactive ^{133}Ba . Experiments were performed in triplicate for 10^{-1} , 1 and 10 mol m^{-3} $BaCl_2$ solutions at 25°C. The D_e value obtained at a $BaCl_2$ concentration of 10 mol m^{-3} , agreed to the estimated value based on the pore diffusion model. The lower $BaCl_2$ concentrations yielded higher D_e values and higher K_d values. The variation in D_e was neither due to the speciation of barium in the solutions nor due to variation in physical properties of the pore structure in the rock because diffusivity of I was the same between 10^{-1} , 1 and 10 mol m^{-3} $BaCl_2$ solutions. Contribution of diffusion in sorbed state should be responsible for the variation in D_e .

Keywords: effective diffusivity, distribution ratio, surface diffusion, granite, barium

稲田花崗岩について Ba^{2*} イオンの透過拡散実験を行い、有効拡散係数(D_e)と分配係数(K_d)を取得した。岩石試料をはさんで 2 つの溶液槽に同じ濃度の安定 $BaCl_2$ 溶液を満たし、放射性の ^{133}Ba を用いて Ba の自己拡散を観察した。溶液として 10^{-1} , 1, 10 mol m^3 $BaCl_2$ 溶液を用い、25 においてそれぞれ 3 ランを行った。10 mol m^3 $BaCl_2$ 溶液を用いた実験で得られた有効拡散係数は細孔拡散モデルから予想される値と一致していた。これに対して塩化バリウムの濃度が低い場合ほど、分配係数が高く、有効拡散係数も高くなった。バリウムの溶存形態は同じであり、同時に拡散させた Γ の挙動から花崗岩試料による間隙構造にも差がないことが明らかになった。それにもかかわらず、有効拡散係数が分配係数に対して正の相関を持っていることは、表面拡散(吸着状態での拡散)の寄与を強く示すものである。

Keywords:有効拡散係数、分配係数、表面拡散、花崗岩、パリウム

1 Introduction

After emplacement of high-level radioactive waste in a deep underground repository, long-lived radionuclides may be leached from the wastes and may subsequently be transported through surrounding rock masses. Major water bearing fractures in rocks surrounding the repository are considered to form main transport paths. Radionuclides diffuse into the pores or micro fissures in the rock matrix and adsorbed on mineral surfaces. These processes lead to retardation of the transport of radionuclides through the fractures. Diffusion of radionuclides through the rock mass is a main transport mechanism under extremely low flow rate of groundwater. To predict the migration of radionuclides in deep geological formations, it is important to understand the diffusion mechanism of radionuclides into the rock matrix and to quantify the diffusivity. Granite was used in this study because it occurs widely in Japan and is considered as a potential host rock for the deep underground disposal of high-level radioactive waste.

Diffusion of aqueous species in macro-porous media has been explained by pore-diffusion model, which was proposed by Brakel & Heertjes [1]. Neretnieks [2] proposed the application of the pore-diffusion model to the diffusion of ions in rock matrix. In the pore-diffusion model, Eq. (1) holds between apparent diffusivity $D_{\rm a}$ (m² s⁻¹), diffusivity in water $D_{\rm v}$ (m² s⁻¹), the constrictivity of the pores δ , the tortuosity of the

pores τ , the bulk density of the rock ρ (kg m⁻³), the distribution ratio $K_{\rm d}$ (m³ kg⁻¹), the porosity of the rock ε , the pore diffusivity $D_{\rm p}$ (m² s⁻¹) and the effective (or intrinsic) diffusivity $D_{\rm e}$ (m² s⁻¹);

$$D_{\rm a} = D_{\rm v} \delta \tau^2 / (1 + \rho K_{\rm d} / \varepsilon) = D_{\rm p} / (1 + \rho K_{\rm d} / \varepsilon) = D_{\rm e} / (\varepsilon + \rho K_{\rm d}). \tag{1}$$
The apparent diffusivity, $D_{\rm a}$, is defined as

$$J=-D_a\partial C/\partial x,$$
 (2)

where J is the diffusive flux (Bq m⁻² s⁻¹), C the amount of a diffusing species in unit volume of porous material (Bq m⁻³) and x the length coordinate in diffusion direction (m). The $D_{\rm e}$ is defined as

$$J=-D_{\circ}\partial c/\partial x$$
. (3)

where c is the concentration of diffusing species in pore water (Bq m⁻³). In the pore-diffusion model, the D_e is given by

$$D_{s} = D_{s} \varepsilon \delta \tau^{2}. \tag{4}$$

Although Eq. (1) and (4) seem to be applicable to describe diffusion of radionuclides in rock matrix, these equations have not been verified. Skagius & Neretnieks [3] determined the D_e of I, tritiated water, Cs and Sr in pieces of granite and found that the results for Cs and Sr were by more than ten times higher than expected from Eq. (4) though the results for non-sorbed species were in fair agreement with Eq. (4). They interpreted the results for Cs and Sr as an effect of surface diffusion. Their determination of the $D_{\rm e}$, however, was not very accurate due to incomplete attainment of the sorption equilibrium. Bradbury & Stephen [4] performed the through-diffusion experiments for $^{85}\mathrm{Sr}$, $^{137}\mathrm{Cs}$ and $^{95\mathrm{m}}\mathrm{Tc}$ in sandstone and found that the D_{e} of Sr was higher than that of I by 4-5 times. They suggested that a different diffusion mechanism or process may be occurring. Bradbury et al. [5] performed the through-diffusion experiments for 85Sr, 137Cs and 95mTc in sandstone anhydrite and upper magnesian limestone and found that the diffusive transport of Sr was higher by orders of magnitude than predicted from the pore-diffusion model. They mentioned that the results may be explained by a second diffusion mechanism, namely surface diffusion. Skagius & Neretnieks [6] performed both the

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in-diffusion experiment and the through-diffusion experiment for Cs and Sr in biotite gneiss and found the diffusive transport was higher than expected from the pore diffusion model. They concluded that both the pore diffusion and the surface diffusion had to be included to interpret the experimental data. Smith [7] performed the through-diffusion experiments for ⁸⁵Sr, ¹³⁷Cs and 125 I in sandstone and found that the $D_{\rm e}$ of 85 Sr is higher than those of ¹³⁷Cs and ¹²⁵I by a factor of more than 6. Tsukamoto & Ohe [8] performed intraparticle diffusion experiments into crushed granite for Cs and Sr and determined the apparent diffusivity. They interpreted the results taking both the surface and the pore diffusion processes. Their determination of the apparent diffusivity, however, was not accurate due to spherical approximation of crushed granite particles in data analysis. Berry & Bond [9] determined the D_e of tritiated water, I, Cs, Sr and Am in sandstone by the through-diffusion method and calculated the contribution of the surface diffusion to the $D_{\rm e}$ as 73 % for Sr and 62 % for Am. They concluded that evidence has been obtained for the apparent existence of the process of surface diffusion in the migration of Sr and Am, but not Cs, through the sandstone. Brace et al. [10] and Ohlsson & Neretnieks [11] suggested the existence of surface diffusion from electrical conductivity measurements, although the evidence is quite indirect.

Several researchers [4-7] observed higher diffusivity for Sr and Cs in rocks than expected from pore diffusion model. However, there was no evidence to show that the higher diffusivity for Sr and Cs is due to the contribution of the surface diffusion.

Our previous study [12] found that the D_e of Sr^{2+} in various rocks roughly agrees with

$$D_{\rm e} = D_{\rm v} \varepsilon \delta \tau^2 + D_{\rm s} \rho K_{\rm d} \tag{5}$$

The second term of Eq. (5), $D_s \rho K_d$, represents the contribution of the surface diffusion. Equation (5) states that the D_e of an ion has the factor that increases linearly with its K_d on rock materials. It is essential to confirm this increase in the D_e with the K_d in order to prove the contribution of the surface diffusion. Systematic through diffusion experiments were performed for Ba²⁺ in a granite as a function of Ba concentration in this study. Experiments were carefully designed to minimize variation in D_e caused by any reason expect for the contribution of the surface diffusion.

2 Experimental

The rock used in this study was a biotitic granite obtained from the Inada mine in the Ibaraki prefecture, eastern Japan. The chemical and mineral compositions were presented elsewhere [8, 12-14]. The porosity and the bulk density of the granite were determined to be 0.49±0.07 % and 2.64x10³ kg m⁻³, respectively [15]. The distribution of the pore diameters was approximately log normal, with a modal diameter of 160 nm [15]. A granite core having a diameter of 40 mm was cut to disks of 5.0 mm thickness with a diamond saw in the same manner as in our previous papers [12, 15]. Granite disks without

visible cracks were used in the experiments. The acrylic diffusion cells used in this study were shown in a previous paper [12]. A granite disk was fitted tightly into the central part of the cell and any gap between the rock disk and the acrylic filled with a silicone gasket. The central support member containing the granite disk was sandwiched between two reservoirs, each with a capacity of 1.16×10^{-4} m³. After assembling a diffusion cell, the diffusion cell was soaked in deionized water under vacuum for 3 days to evacuate all air from the interconnected pores.

After the evacuation was completed, the diffusion cell was filled with a 10 mol m⁻³ BaCl₂ solution to pre-equilibrate the granite disks with the solution for 30 days. The starting solution was prepared by combining 1.20x10⁻⁴ m³ of 10 mol m⁻³ BaCl₂ solution, $3.0x10^{-7}$ m³ of ¹³³Ba (10.5 y half-life) stock solution $(10^3 \text{ mol m}^{-3} \text{ HCl})$ and $3.0 \times 10^{-7} \text{ m}^3 \text{ of } ^{125} \text{I}$ (59.9 d half-life) stock solution in a polypropylene bottle. The stock solution of ¹²⁵I was a NaI/Na₂S₂O₃ mixture solution in which the concentrations of Γ and $S_2O_3^{2-}$ were $3.3x10^{-1}$ and $3.2x10^{-1}$ mol m⁻³, respectively. The concentrations of ¹³³Ba and ¹²⁵I in the starting solution were planned to be 1040 MBq m⁻³ and 2070 MBq m⁻³, respectively. To measure the concentrations of ¹³³Ba and ¹²⁵I in the solutions, a 1.0x10⁻⁶ m³ aliquot was withdrawn and its activity determined by γ-spectrometry. The statistical error of the determination of the tracer concentration was 1 -10 % by radioactivity measurement. The solutions were prepared from reagent grade chemicals (Wako Pure Chemical Industries, Ltd., Tokyo) and deionized water (Milli-Q Labo System, Millipore). The diffusion experiment was started by placing the starting solution in the source reservoir and a blank 10 mol m⁻³ BaCl₂ solution in the other, or measurement reservoir. The diffusion experiment was performed at (25.0±0.5)°C in a water bath. The diffusion experiment for the 10 mol m⁻³ BaCl₂ solution was performed in triplicate using 3 granite samples, L14, N9 and O3. The name of the granite sample, L14, denotes the 14th disk cut from the granite core named L.

The diffusion experiment was performed in triplicate also for 1 mol m⁻³ BaCl₂ solution using 3 granite samples, N6, O6 and O8, and for 10⁻¹ mol m⁻³ BaCl₂ solution, granite samples, L12, O16 and O22.

At 14-day intervals, a 1.0×10^{-6} m³ aliquot was taken from the measurement reservoirs to determine the concentrations of 133 Ba and 125 I. The 1.0×10^{-6} m³ aliquot removed from the measurement reservoir was replaced with an equal volume of the BaCl₂ solution to maintain the balance of the water level between the two reservoirs. Balancing is necessary to avoid development of a pressure gradient that could lead to advective transport from the source to the measurement reservoir across the granite disk. The concentrations of 133 Ba and 125 I in the source reservoir were determined occasionally by taking a 1.0×10^{-8} m³ aliquot. At the termination of each run, the inner wall of the measurement reservoir was rinsed with 10^3 mol m⁻³ HCl to determine the amount of 133 Ba adsorbed on the cell walls. The amount of 133 Ba was found to be less than 1 % of the final

calculation.

¹³³Ba inventory in the measurement reservoir and can be ignored

Zeta potential of the granite in $BaCl_2$ solutions were measured by a microscope electrophoresis zeta potential analyzer, RANK Brothers MARK II. Crushed granite material (2 mm) was milled by a mixer mill. The milled material was suspended in the $BaCl_2$ solutions overnight and the supernatant was used for the zeta potential analysis.

3 Results and discussion

Figure 1(a) shows the time dependence of the concentrations of 133 Ba in the measurement reservoirs from the diffusion experiment using the 10 mol m⁻³ BaCl₂ solution. The concentrations of 133 Ba were corrected for the decay. The concentrations increase linearly after 100 days from the initiation of the diffusion. The rate of change in concentration of 133 Ba due to diffusion in pore water at distance x from the surface facing to the source reservoir at time t can be described as

$$(\varepsilon + \rho K_{\rm d})\partial c/\partial t = D_{\rm e}\partial^2 c/\partial x^2 \tag{6}$$

This equation is based on an assumption that the $K_{\rm d}$ is independent on t and on x. Skagius & Neretnieks [6] presented the data analysis for the case that $K_{\rm d}$ depends on t and on x assuming non-linear sorption isotherms. In this study, a BaCl₂ solution of the same concentration was placed in both reservoirs and was pre-equilibrated with the rock disk before starting diffusion of 133 Ba. This assures that the concentration of barium is independent on t and on x, and that the $K_{\rm d}$ can be assumed to be independent on t and on x. The initial and boundary conditions are

$$c(x, 0) = 0 \text{ at } 0 < x \quad L$$
 (7)

$$c(0,t) = c_1 \tag{8}$$

$$c(L, t) = c_2 << c_1,$$
 (9)

where L is the thickness of the rock sample $(5x10^{-3} \text{ m})$, c_1 the concentration of 133 Ba in the source reservoir (Bq m⁻³), c_2 the concentration of 133 Ba in the measurement reservoir (Bq m⁻³). In the case that c_1 can be regarded to be constant, the solution of Eqs. (6) - (9) was given by Crank [16] and the concentration of 133 Ba in the measurement reservoir after a long period is approximated as

$$c_2(t) = AV_2^{-1}c_1(D_eL^-1t - (\varepsilon + \rho K_d)L/6)$$
 (10) where A is the cross section of the sample $(1.256 \times 10^{-3} \text{ m}^2)$, V_2 the volume of the measurement reservoir $(1.16 \times 10^{-4} \text{ m}^3)$. The least-squares fitting of the plots of $c_2(t)$ vs. t to Eq. (10) as shown in Fig. 1(a) yields the D_e from the slope, and the rock capacity factor, $(\varepsilon + \rho K_d)$, from the intercept on the concentration axis of the extrapolated linear portion. The D_e of $(1.95 \pm 0.21) \times 10^{-13} \text{ m}^2 \text{s}^{-1}$ and the K_d of $(8.0 \pm 0.9) \times 10^{-5} \text{ m}^3 \text{ kg}^{-1}$ were obtained on an average of three runs and tabulated in

Table 1. An average ¹³³Ba concentration in the source reservoir

in the period of 100 - 400 days was used as c_1 in this

Figure 1(b) shows the results from the diffusion experiment using the 1mol m⁻³ BaCl₂ solution. The linear portion of the curve appeared after 200 days from the initiation of the diffusion. The $D_{\rm e}$ of $(2.38\pm0.38) \times 10^{-13}$ m² s⁻¹ and the $K_{\rm d}$ of $(2.26\pm0.35) \times 10^{-4}$ m³ kg⁻¹ were obtained. An average ¹³³Ba concentration in the source reservoir in the period of 200 – 580 day was used as c_1 in this calculation.

Figure 1(c) shows the results from the diffusion experiment using the 10^{-1} mol m⁻³ BaCl₂ solution. The linear portion of the curve appeared after 300 days from the initiation of the diffusion. The $D_{\rm e}$ of $(6.6\pm0.7) \times 10^{-13}$ m² s⁻¹ and the $K_{\rm d}$ of $(8.4\pm0.9) \times 10^{-4}$ m³ kg⁻¹ were obtained. An average 133 Ba

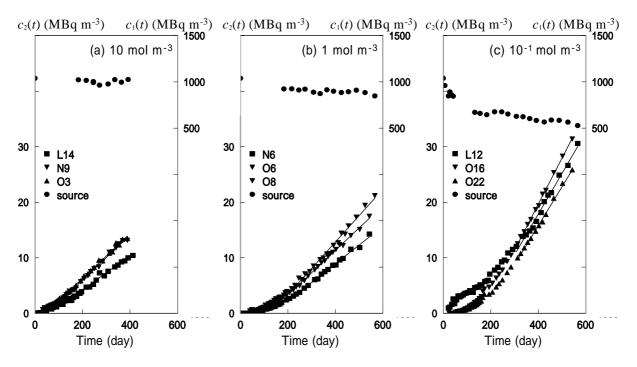


Fig. 1 Changes in concentration of 133 Ba in measurement reservoir due to diffusion through 5-mm thick Inada granite samples obtained in 10 mol m $^{-3}$ BaCl₂ solution (a), 1 mol m $^{-3}$ BaCl₂ solution (b), 10^{-1} mol m $^{-3}$ BaCl₂ solution (c).

Table 1 The D_e and the K_d of Ba in granite.

[BaCl ₂] (mol m ⁻³)	Sample No.	$\begin{array}{c} D_{\rm e} \\ (10^{\text{-}13} \text{ m}^2 \text{ s}^{\text{-}1}) \end{array}$	$(10^{-4} \mathrm{m}^3 \mathrm{kg}^{-1})$
10-1	L12	6.57±0.21	8.1±0.7
	O16	7.43 ± 0.09	9.56±0.30
	O22	5.71 ± 0.11	7.59 ± 0.34
	Avg.	6.6 ± 0.7	8.4 ± 0.9
1	N6	1.99±0.04	1.94±0.10
	O6	2.25 ± 0.03	2.12 ± 0.09
	O8	2.90 ± 0.07	2.72 ± 0.20
	Avg.	2.38 ± 0.38	2.26 ± 0.35
10	L14	1.66±0.03	0.74±0.06
	N9	2.12 ± 0.03	0.90 ± 0.05
	O3	2.08 ± 0.03	0.76 ± 0.07
	Avg.	1.95±0.21	0.80 ± 0.09

concentration in the source reservoir in the period of 300 - 580 days was used as c_1 in this calculation.

Figure 2(a) shows the time dependence of the concentration of ^{125}I in the measurement reservoirs from the diffusion experiment using the 10 mol m⁻³ BaCl₂ solution. The concentrations of ^{125}I were corrected for the decay. The concentrations increase linearly for the first 70 days period. The increase in the concentration of ^{125}I become less steep in the period of 70 – 120 days and another linear increase was observed in the period of 120 - 250 days. From the first linear portion of the curve, a $D_{\rm e}$ of $(6.1\pm0.9)\text{x}10^{-13}$ m² s⁻¹ and a $K_{\rm d}$ of $<1.6\text{x}10^{-5}$ m³ kg⁻¹ were obtained on an average of three runs and tabulated in Table 2. The second linear portion was analyzed using Eq. (3) to obtain another $D_{\rm e}$ as

 $D_{\rm e} = -J/(\partial c/\partial x) = (\Delta c_2/\Delta t)VA^{-1}/\{(\bar{\rm c}_1 - \bar{\rm c}_2)L^{-1}\}$ (11) where $\Delta c_2/\Delta t$ is the rate of increase in the concentration of ¹²⁵I in the period between 120 and 250 days (Bq m⁻³ s⁻¹), $\bar{\rm c}_1$ the average concentration of ¹²⁵I in the source reservoir in the

Table 2 The D_e and the K_d of I in granite.

in the first of th							
[BaCl ₂] Sample No.		0 – 70 day		120 - 250			
(mol m ⁻³)				day			
		D_{e}	K_{d}	$D_{ m e}$			
		$(10^{-13} \text{ m}^2 \text{ s}^{-1})$	$(10^{-4} \mathrm{m}^3 \mathrm{kg}^{-1})$	$(10^{-13} \text{ m}^2 \text{ s}^{-1})$			
10-1	L12	10.05±0.48	0.55±0.19	3.06 ± 0.24			
	O16	3.90 ± 0.47	-0.53 ± 0.18	3.54 ± 0.29			
	O22	6.29 ± 0.39	-0.38 ± 0.14	3.89 ± 0.23			
	Avg.	6.7 ± 2.6	< 0.39	3.50 ± 0.42			
1	N6	6.62±0.36	0.09±0.10	2.20±0.31			
	O6	8.55 ± 0.94	-0.27 ± 0.32	1.41 ± 0.31			
	O8	10.08 ± 0.31	0.32 ± 0.09	4.74 ± 0.47			
	Avg.	8.4 ± 1.7	< 0.26	2.8 ± 1.5			
10	L14	5.15±0.10	-0.12±0.03	3.18±0.17			
	N9	7.22 ± 0.31	-0.03 ± 0.10	3.07 ± 0.21			
	O3	6.07 ± 0.13	-0.13 ± 0.04	3.91 ± 0.21			
	Avg.	6.1 ± 0.9	< 0.16	3.39 ± 0.43			

period of 120 - 250 days, \bar{c}_2 the average concentration of ^{125}I in the measurement reservoir in the period of 120 - 250 days. A D_e of $(3.39\pm0.43) \times 10^{-13}$ m² s⁻¹ was obtained by the least square fitting and tabulated in Table 2. Figure 2(b) shows the results from the diffusion experiment using the 1 mol m⁻³ BaCl₂ solution. From the first linear portion of the curve, a D_e of $(8.4\pm1.7) \times 10^{-13}$ m² s⁻¹ and a K_d of <2.6 $\times 10^{-5}$ m³ kg⁻¹ were obtained. From the second linear portion of the curve, another D_e of $(2.8\pm1.5) \times 10^{-13}$ m² s⁻¹ was obtained. Figure 2(c) shows the results from the diffusion experiment using the 10^{-1} mol m⁻³ BaCl₂ solution. From the first linear portion of the curve, a D_e of $(6.7\pm2.6) \times 10^{-13}$ m² s⁻¹ and a K_d of <3.9×10⁻⁵ m³ kg⁻¹ were obtained. From the second linear portion of the curve, another D_e of $(3.50\pm0.42) \times 10^{-13}$ m² s⁻¹ was obtained.

The $D_{\rm e}$ of I was reduced by half about 100 days after the start of the diffusion as shown in Table 2. This reduction may be due to oxidation of Γ to ${\rm IO_3}^-$. In aerated conditions, Γ can be oxidized by dissolved oxygen;

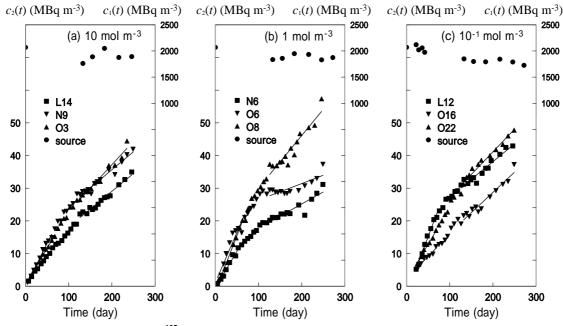


Fig. 2 Changes in concentration of ^{125}I in measurement reservoir due to diffusion through 5-mm thick Inada granite samples obtained in 10 mol m $^{-3}$ BaCl₂ solution (a), 1 mol m $^{-3}$ BaCl₂ solution (b), 10^{-1} mol m $^{-3}$ BaCl₂ solution (c).

$$I^- + 1.5O_2(g) = IO_3^- log K = 13.38 [17]$$
 (12)
Diffusivity of IO_3^- in water at infinite dilution is $1.06 \times 10^{-9} \text{ m}^2 \text{ s}^{-1}$, which is about half as high as that of $I^- (2.00 \times 10^{-9} \text{ m}^2 \text{ s}^{-1})$ at 25°C [18].

The pore-diffusion model gives the $D_{\rm e}$ as Eq. (4) where $\varepsilon \delta \tau^2$ is $(2.6\pm0.4) \times 10^{-4}$ [12] for the granite. Since the diffusion of ¹³³Ba in the present experiment was self - diffusion, diffusivity of Ba^{2+} at infinite dilution, $8.48 \times 10^{-10} \mathrm{m}^2 \mathrm{s}^{-1}$ [18], can be used as D_{v} . Equation (4) gave the D_e of Ba²⁺ of (2.2 ± 0.3) x 10^{-13} m² s⁻¹ under the present experimental conditions. The $D_{\rm e}$ values obtained at the BaCl₂ concentration of 10 mol m⁻³ agreed with the estimated value. The D_e value obtained for the 1 mol m⁻³ BaCl₂ solution was slightly higher than the value and the value obtained for the 10⁻¹ mol m⁻³ BaCl₂ solution was higher by 3 times. It is noticeable that if the decrease in the concentration of ¹³³Ba in the source reservoir shown in Fig. 1(c) was taken into account by using the method developed by Spacek & Kubin [19], the D_e of 1.3×10^{-12} m² s⁻¹ was obtained, which is higher by 6 times than the value estimated by the pore-diffusion model. But the D_e value obtained by their method was not further considered in this paper because comparison with the other $D_{\rm e}$ values obtained by Eq. (10) is not convincing.

The K_d of barium on granite was lowest in case that the concentration of BaCl₂ was 10 mol m⁻³ and was highest in case of 10^{-1} mol m⁻³. This type of dependence of the K_d on concentration is often observed for Sr that analogizes to barium [20-22], and can be attributed to the gradual saturation of adsorption sites on granite by Ba²⁺. The zeta potential of the milled granite suspension was -30.7 mV in 10^{-1} mol m⁻³ BaCl₂ solution, -13.6 mV in 1 mol m⁻³ BaCl₂ solution and -7.4 mV in 10 mol m⁻³ BaCl₂ solution. The fact that the zeta potential of the granite gradually approached zero as the BaCl₂ concentration increased suggested that the sorption sites were gradually dominated by Ba²⁺ ions at the higher concentration of BaCl₂.

In case of low $K_{\rm d}$ value due to high BaCl₂ concentration, the $D_{\rm e}$ value was low, while in case of higher $K_{\rm d}$ value due to low BaCl₂ concentration, the $D_{\rm e}$ value was high. The positive correlation is clearly seen between the $D_{\rm e}$ and the $K_{\rm d}$ as the second term of Eq. (5) stated.

The variation in the D_e value in the three BaCl₂ concentrations was not due to the speciation of barium in the solution because complexation of Ba²⁺ by Cl⁻ is negligible and carbonate precipitation is not favorable under the working pH range (5.8 - 6.7) [23].

The $D_{\rm e}$ of Γ , non-sorbed species, was not affected by BaCl₂ concentrations; $(6.1\pm0.9) \times 10^{-13}$, $(8.4\pm1.7) \times 10^{-13}$ and $(6.7\pm2.6) \times 10^{-13}$ m² s⁻¹ for 10, 1 and 10^{-1} mol m⁻³ BaCl₂ solution, respectively. These values are close to an estimated value, 6.4×10^{-13} m² s⁻¹, obtained from an equation for the $D_{\rm e}$ of Γ in rocks [24]

$$D_{\rm e} = 6.4 \times 10^{-10} \varepsilon^{1.3},\tag{13}$$

where ϵ was 0.0049. The similar D_e values of Γ for the three BaCl₂ concentrations indicate that the physical property of the pore structure was almost the same between the experiments, and accordingly the variation in the D_e of Ba²⁺ was not

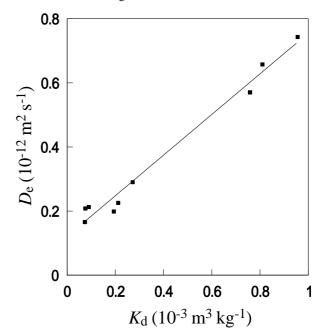


Fig. 3 Correlation between the $D_{\rm e}$ and the $K_{\rm d}$ of ¹³³Ba in the granite.

physically caused. Both $D_{\rm v}$ and $\varepsilon \delta \tau^2$ values were independent on the concentration of BaCl₂; nevertheless, the $D_{\rm e}$ of Ba²⁺ depended on concentration of BaCl₂. The second term of Eq. (5) or the contribution of the surface diffusion should be responsible for the variation in the $D_{\rm e}$.

The $D_{\rm e}$ of Ba²⁺ was plotted versus $K_{\rm d}$ in Fig. 3. Using the least squares method, the $D_{\rm s}$ value was determined to be $(2.4\pm0.1){\rm x}10^{-13}~{\rm m}^2~{\rm s}^{-1}$. The present $D_{\rm s}$ value is smaller than previously obtained value, $3.5{\rm x}10^{-12}~{\rm m}^2~{\rm s}^{-1}$, for Sr²⁺ [12].

4 Conclusion

Through diffusion experiments for barium ion in granite were performed; a positive correlation was found between the $D_{\rm e}$ and the $K_{\rm d}$. The result is strongly indicative of the diffusion in sorbed state. When we apply pore diffusion model to sorbing ions, the diffusive transport can be underestimated.

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